Research Brief

Summer methane ebullition from a headwater catchment in Northeastern Siberia

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Abstract

Streams and rivers are active processors of terrestrial carbon and significant sources of carbon dioxide (CO_2) and methane (CH_4) to the atmosphere. Recent studies suggest that ebullition may represent a sizable yet overlooked component of the total CH_4 flux from these systems; however, there are no published CH_4 ebullition estimates for streams or rivers in subarctic or arctic biomes, regions that store vast quantities of vulnerable, old organic carbon in permafrost soils. We quantified CH_4 ebullition from headwater streams in a small arctic watershed in Northeastern Siberia. Ebullitive emissions were 0.64 mmol m^{-2} d^{-1}, which is lower than the global average but approximately 2 times greater than the pan-arctic diffusive CH_4 flux estimate reported in a recent synthesis of global freshwater CH_4 emissions. The high CO_2:CH_4 of sediment bubbles (0.52) suggests that methane emissions may currently be constrained by resource competition between methanogens and microbes using more efficient metabolic strategies. Furthermore, the magnitude and frequency of ebullition events were greater as temperatures increased, suggesting that ebullition from streams could become a more prominent component of the regional CH_4 flux in a warmer future.

Key words: Arctic, carbon dioxide, carbon emissions, ebullition, methane, rivers, Siberia, streams

Introduction

Rivers and streams are important sites of terrestrial carbon (C) transport and processing (Cole et al. 2007, Battin et al. 2009) and have gained recognition as considerable sources of carbon dioxide (CO_2) to the atmosphere (Aufdenkampe et al. 2011, Butman and Raymond 2011, Raymond et al. 2013). Their role in the global methane (CH_4) cycle, however, remains uncertain, particularly in high-latitude environments (Bastviken et al. 2011) where permafrost soils store vast quantities of old organic C (Tarnocai et al. 2009, Hugelius et al. 2014). As the climate continues to warm and permafrost thaws, this C will be increasingly vulnerable to mineralization and transfer to the atmosphere as greenhouse gases, thereby perpetuating a potent positive feedback to anthropogenic climate warming (Schuur et al. 2008, 2013). Understanding the

factors regulating C metabolism in streams and rivers and the relative contribution of these systems to aquatic CH_4 emissions is therefore necessary to constrain flux estimates and predict future emissions from arctic and subarctic regions (Kirschke et al. 2013).

Gas flux estimates from streams and rivers typically emphasize diffusive emissions and overlook the contribution of ebullition (bubble mediated emission), which is the dominant pathway of CH_4 evasion from lakes (Bastviken et al. 2004). Recent studies in temperate and tropical streams and rivers suggest that ebullition may represent up to 80% of CH_4 emissions from these systems, but data remain sparse (Baulch et al. 2011, Sawakuchi et al. 2014). Ebullition is both episodic and spatially variable and is often poorly quantified due to insufficient spatial coverage and/or brief measurement periods (Wik et al. 2013, Maeck et al. 2014). In lakes, the timing and magnitude of

ebullition events is often driven by landscape-level, meteorological factors such as falling hydrostatic pressure (Mattson and Likens 1990, Varadharajan and Hemond 2012) and energy inputs (Wik et al. 2014), but the factors driving these events in streams remain uncertain. Here we report the first estimate of CH_4 ebullition from streams north of 49°N and explore possible factors affecting the spatial and temporal patterns of ebullitive emissions.

Study site

We studied stream reaches within the "Y3" catchment of the lower Kolyma River basin (Fig. 1) located near the Northeast Science Station in the town of Cherskiy, Russia. The Kolyma River is the fifth largest river draining into the Arctic Ocean, and its watershed is the world's largest underlain entirely by continuous permafrost. The Y3 watershed covers 16.84 km^2 and extends from alpine to floodplain ecosystems (elevation: 18–275 m). The regional climate is dry, cold, and continental. Sampled streams drain boreal larch (*Larix cajanderi*) forests of varying density and large patches of dense, low-lying *Betula*, *Salix*, and *Vaccinium* shrubs. The watershed also contains 9 small lakes, one of which was the subject of 2 intensive studies of CH_4 ebullition (Zimov et al. 1997, Walter et al. 2006).

Methods

We measured CH_4 and CO_2 ebullition from streams of the Y3 watershed for 40 d during July and August 2014 using inverted funnels affixed to wooden stakes to trap bubbles (Molongoski and Klug 1980). We deployed 24 funnel traps at sites chosen randomly along 6 accessible stream reaches (order 2–4) using a Geographic Information System (GIS). Accumulated gas volumes were recorded every 2–4 d. On 25 July and 13 August we collected sediment bubbles from sites adjacent to each trap (within 1 m) to analyze CO_2 and CH_4 content. Bubble samples were collected by stirring sediments to release bubbles and capturing them in a mobile version of our stationary bubble traps. Stirred bubbles were collected rather than those collected in our traps due to the potential for diffusion of component gases out of the trap (Crawford et al. 2014). Due to the shallow water depth of our streams (<1 m), we assumed that equilibration of erupted bubbles with the surrounding water column (McGinnis et al, 2006) was negligible and that stirred bubbles were representative of the true ebullitive flux (Crawford et al. 2014).

At each site, we collected one 10 mL gas sample and transferred it into a 24 mL serum vial pre-flushed with N_2 gas and sealed with thick butyl rubber stoppers. Samples were shipped to Fort Collins, Colorado, for analysis on a
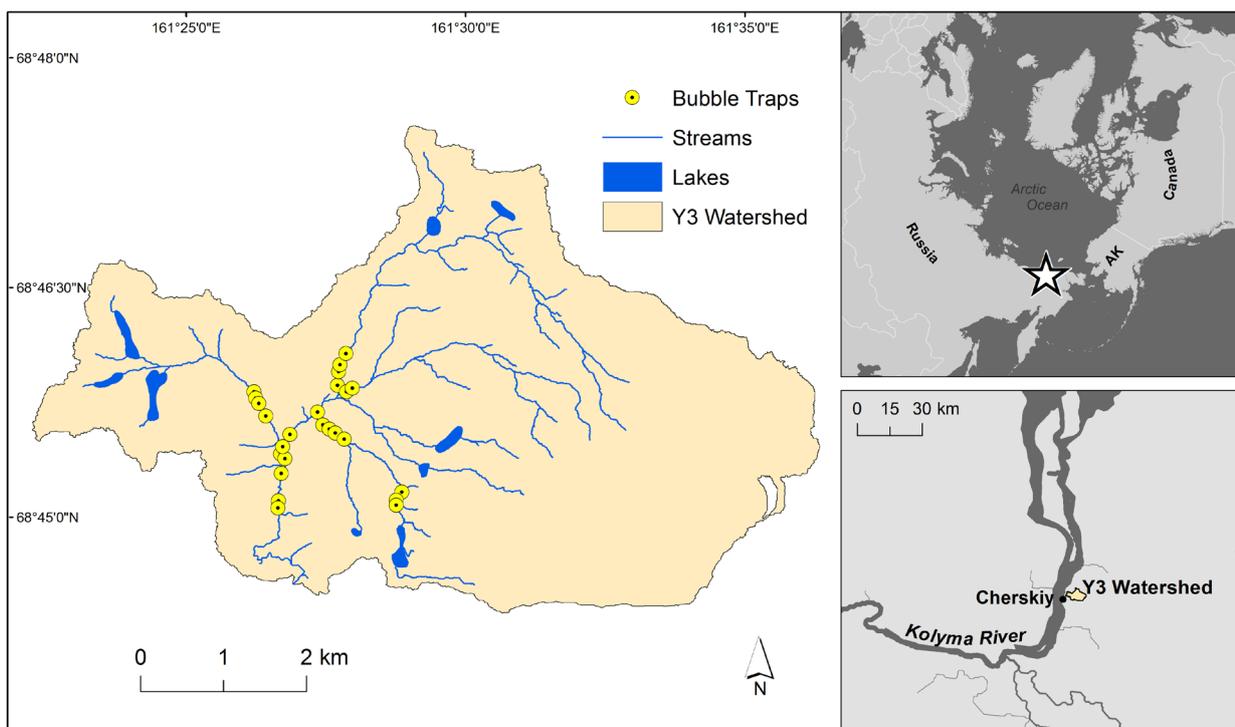
Figure 1. Location of our study site, the Y3 watershed, and the spatial distribution of bubble traps.

Table 1. Y3 sediment metrics for July and August 2014. Data represent the means of all sites (± standard error). See Supplementary Table S1 for site-specific sediment metrics.

Month	Bulk Density (g cm⁻³)	OM Content (%LOI)	Thaw Depth (cm)	Temperature (°C)
July	—	—	56.3 (±2.3)	5.3 (±0.3)
August	0.9 (±0.1)	17.0 (±3.4)	57.0 (±2.6)	8.1 (±0.5)

Los Gatos Research (Mountain View, CA) DLT-100 greenhouse gas analyzer (see Supplementary Methods). Prepared standards accompanied samples in shipment to evaluate potential leakage during transit. After accounting for the dilution resulting from our sampling procedure, we estimated daily molar fluxes of CH_4 and CO_2 according to the ideal gas law as:

$$n = (P \times V) / (R \times T) \times X_{gas}, \qquad (1)$$

where n is the molar flux of a given gas, P is barometric pressure (atm), V is the rate of bubble release (L m⁻² d⁻¹), R is the gas constant (L atm K⁻¹ mol⁻¹), T is the sediment temperature at the time of bubble collection (K), and X_{gas} is the mole fraction of CO_2 or CH_4 in collected bubbles.

We used Monte Carlo methods to calculate a constrained estimate of total ebullitive CO_2 and CH_4 fluxes from Y3 streams with quantified uncertainty (Baulch et al. 2011, Crawford et al. 2013, 2014). We created probability distributions of CO_2 and CH_4 fluxes by resampling (n = 10 000) measured flux volumes, bubble compositions, and a probability distribution of stream area. The stream area distribution was created using Monte Carlo resampling as the product of stream length and a random survey of 96 stream widths measured throughout Y3 in late July 2014. Stream length was generated using a digital elevation model (DEM) derived from high-resolution (<2 m) stereo-pair imagery obtained from the Polar Geospatial Center (www.pgc.umn.edu) and validated by field observation. We assumed a 120 d open water season. Fluxes were reported as the mean of the resulting distribution, and error was quantified as a 95% confidence interval (CI).

To explore factors affecting ebullition events, we collected 3 cm diameter cores from the upper 10 cm of sediment at each site in early August to measure bulk density and organic matter (OM) content. Sediment OM was determined as percent loss on ignition (%LOI) by combusting samples at 450 °C for 4 h in a muffle furnace. We also measured sediment temperature using a type K thermocouple and thaw depth (distance between sediment surface and underlying permafrost) with a metal probe at each site, once monthly. To identify potential meteorological controls, we assessed the temporal consistency of ebullition events between sites by correlation, assuming

that these landscape-level controls should affect all sites comparably, and fluxes between sites should therefore be positively correlated (Crawford et al. 2014).

Data distributions were described using Fisher-Pearson's coefficient of skewness ($\gamma 1$) and the coefficient of excess kurtosis (γ_2). Because our data were not normally distributed, significant differences in monthly bubble composition and sediment temperature were identified with a Wilcoxon signed-rank test. Relationships between sediment metrics and mean flux magnitude, cumulative flux, and bubble composition were explored using linear regression. All data were log-transformed to meet the assumptions of linear regression, as determined and confirmed using the *gvlma* package in R (Pena and Slate 2014). The sample size (n) for all statistical analyses was 24, unless otherwise noted.

Results

Volumetric bubble release rates were spatially and temporally variable (Fig. 2), ranging from 0 to 2420 mL m⁻² d⁻¹ with a mean of 83 mL m⁻² d⁻¹. The distribution was skewed right with high kurtosis ($\gamma_1 = 6.26$, $\gamma_2 = 42.6$). Flux events were episodic, and the timing of these events was not consistent among sites, with most sites showing both positive and negative correlations with other sites (Supplementary Fig. S1). In general however, we observed an increase in the relative frequency and magnitude of bubble flux at 19 of 24 sites beginning in late July (Supplementary Fig. S2). Neither mean flux rates nor cumulative flux volumes were correlated with sediment bulk density, OM content, temperature, or thaw depth (p > 0.05, $R^2 < 0.1$).

The mean CH_4 mole fraction of sediment bubbles was 23.3% (range: 1.7–98%) and the distribution was highly skewed to the right ($\gamma_1 = 1.43$, $\gamma_2 = 1.73$). On average, CO_2 composed an additional 4.0% (range: 1.0–11.4%) of bubble component gas, and the distribution was also right-skewed ($\gamma_1 = 1.03$, $\gamma_2 = 0.36$). Mean bubble CO_2:CH_4 was 0.52 (range: 0.02–2.48). While both bubble CO_2 content and sediment temperature were significantly higher in August compared to July (p < 0.001; Fig. 3), neither bubble CH_4 content nor CO_2:CH_4 showed significant changes between July and August (p > 0.05). Bubble composition was not correlated to sediment bulk density, OM content, or thaw depth (Table 1; p > 0.05, $R^2 < 0.1$).

Ebullitive CH_4 flux from Y3 streams was 0.64 mmol m^{-2} d^{-1} (95% CI: 0–5.82 mmol m^{-2} d^{-1}), and CO_2 was 0.13 mmol m^{-2} d^{-1} (95% CI: 0–1.14 mmol m^{-2} d^{-1}). Median surface area of Y3 streams was 0.135 km^2 (95% CI: 0.022–0.869 km^2), which represents ~0.8% of the total watershed area. Scaling our fluxes to this area, we estimate that ebullition from streams contributes an additional 40 kg CO_2-C (95% CI: 0–315 kg) and 186 kg CH_4-C (95% CI: 0–1.51 Mg) to aquatic C emissions from the Y3 watershed during the open water season.

Discussion

Our study is the first to quantify ebullition from streams in the Arctic and is one of few stream ebullition studies conducted anywhere in the world. Mean ebullitive CH_4 flux from Y3 is lower than most other studies and 35% lower than the global average (Table 2). Ebullition from

streams in this study represented ~13% of total aquatic CH_4 ebullition from the Y3 watershed, an area encompassing lakes that represent the largest lake CH_4 emissions in the literature by 1–3 orders of magnitude (Walter et al. 2006, Bastviken et al. 2011). While CH_4 ebullition from this study was lower than most other regions, CO_2 flux estimates were comparable to those of a temperate peatland stream (Crawford et al. 2014). Compared to diffusive emissions reported for Y3 and neighboring streams (Denfeld et al. 2013), however, CO_2 ebullition was negligible, constituting <0.0002% of the total CO_2 flux from streams.

The quantity of CH_4 in bubbles and the magnitude of CH_4 emissions may currently be constrained by sediment redox conditions. The high CO_2:CH_4 of Y3 sediment bubbles compared to streams in other regions (Table 2) suggests that arctic stream sediments may be less reduced than in temperate and tropical streams (Schimel 1995).
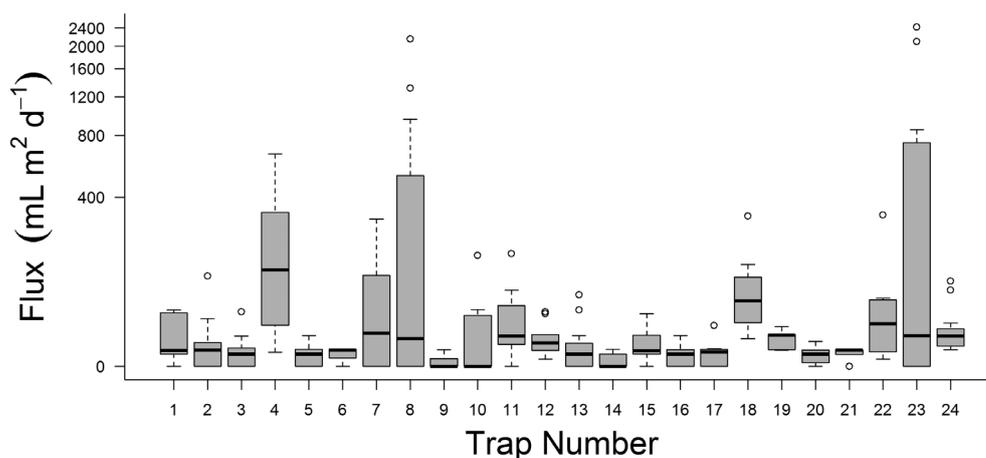
Figure 2. Boxplots showing the flux volume distribution for each site. Thick, horizontal black lines represent median flux volumes; boxes represent the interquartile range (IQR); fences indicate 1.5×IQR; and open circles represent outliers.
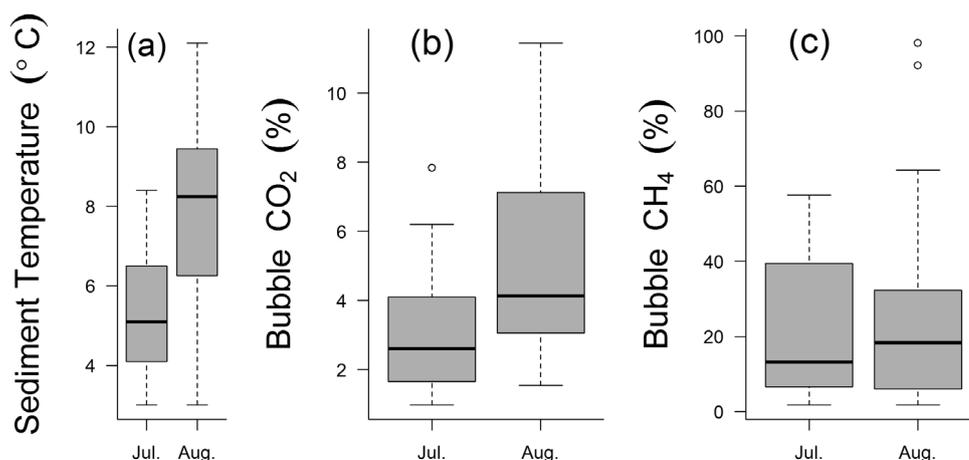
Figure 3. Boxplots of (a) monthly sediment temperature at 15 cm depth, (b) monthly bubble CO_2 content, and (c) monthly CH_4 concentration. Thick, horizontal black lines represent median flux volumes; boxes represent the interquartile range (IQR); fences indicate 1.5×IQR; and open circles represent outliers.

Table 2. Comparison of ebullitive CH_4 emissions from streams and rivers of diverse ecoregions organized by latitude. Error is reported as standard error for our study (n = 48 for %CH_4 and CO_2:CH_4; n = 238 for Mean Volume and Mean CH_4 Flux).

Region	Study	Latitude	Mean Volume (mL m^{-2} d^{-1})	Mean Mole Fraction CH_4 (%)	CO_2:CH_4	Mean Flux (mmol m^{-2} d^{-1})
Russian Arctic	This Study	69°N	83 (±18)	23.3(±3.3)	0.52 (±0.8)	0.64 (±0.13)
Temperate Germany	Maeck et al. 2013	49°N	—	—	—	0.04
Temperate USA	Crawford et al. 2014	46°N	123	22	0.08	1.25
Temperate Canada	Baulch et al. 2011	44°N	104	26	0.01– 0.07[a]	1.41
Amazon River and Tributaries	Sawakuchi et al. 2014	2°S	—	—	—	0.6
Agricultural Streams, New Zealand	Wilcock and Sorrell 2008	37°S	—	41	0.03–0.3[a]	1.68
Global Average						0.94

[a] Range calculated as the minimum and maximum ratios from the range of CO_2 and CH_4 mole fractions reported.
"—" Indicates that data were not reported.

Accumulations of oxidized iron, Fe(III), which were common throughout reaches of Y3 during this study, lend qualitative support to this hypothesis because the presence of Fe(III) and alternative electron acceptors can inhibit methanogenesis in anoxic sediments (Roden and Wetzel 1996, Bodegom et al. 2004). These metabolic strategies produce CO_2, which is relatively water soluble and does not readily nucleate bubbles (van Kesteren and van Kessel 2002). If sediment respiration was dominated by anaerobic CO_2 production, we would expect to see bubbles with both low CO_2 and CH_4 content and high CO_2:CH_4, as observed.

Sediment temperature may also regulate the rate of bubble production and emission. While barometric pressure is a driver of ebullition events in lakes (Mattson and Likens 1990, Varadharajan and Hemond 2012), variation in the timing of bubbling events across sites suggests that pressure changes were not the primary driver of bubbling in these streams during the time period of our study. Instead, we hypothesize that the magnitude and frequency of ebullition events depend on the rate of bubble production and growth within sediments. Bubble production rate is related to both respiration rate and the solubility of component gases (Fendinger et al. 1992, Boudreau 2012) and is therefore temperature dependent (Yvon-Durocher et al. 2012, 2014). While the size of our dataset and the length of our study do not permit us to quantitatively assess the role of temperature on bubble release rates in Y3, observed late-season increases in flux volume and frequency (Supplementary Fig. S2) corresponded to a period of warmer air, water, and sediment temperature. Noting the temperature dependence of CO_2 production in Y3 (Fig. 3), we suspect that this late-season increase is due to greater rates of sediment respiration and bubble production following ice-out and a sufficient sediment warm-up period (Maeck et al. 2013). Energy input, through its effect on sediment temperature and substrate availability, is a known driver of ebullition in lakes (Wik et al. 2014). We surmise that this may also be true in streams and recommend that future studies consider localized variation in energy flux due to riparian shading (Bott et al. 2006, Ball et al. 2010, Burrell et al. 2014) as a potential source of flux variability.

Ebullitive CH_4 emissions from streams are both regionally significant and sensitive to environmental change. In a recent global synthesis of freshwater CH_4 emissions (Bastviken et al. 2011), high-latitude stream emissions were extrapolated from a single study of diffusive emissions from the Kuparuk River in Northern Alaska (Kling et al. 1992). The flux estimate from this river (0.36 mmol m^{-2} d^{-1}) only accounted for diffusion and is ~2 times less than the emissions we observed via ebullition, suggesting that ebullition could be a considerable component of CH_4 emissions from arctic streams and that the current pan-arctic estimate of CH_4 emissions from rivers and streams may be considerably underestimated. In addition, more recent diffusive estimates from boreal, subarctic, and arctic streams and rivers range from 0.63 to 15.8 mmol m^{-2} d^{-1} (Striegl et al. 2012, Crawford et al. 2013, Huotari et al. 2013, Lundin et al. 2013), lending further support to the importance of CH_4 emissions from

arctic streams to the regional and global CH$_4$ cycle. The Arctic is expected to continue warming throughout the next century at rates greater than the global average (Overland et al. 2014). If rates of ebullition are sensitive to temperature, as we hypothesize, the role of CH$_4$ ebullition from arctic streams may become more prominent in a warmer future.

Acknowledgements

References

Aufdenkampe AK, Mayorga E, Raymond PA, Melack JM, Doney SC, Alin SR, Aalto RE, Yoo K. 2011. Riverine coupling of biogeochemical cycles between land, oceans, and atmosphere. Front Ecol Environ. 9:53–60.

Ball B, Kominoski J, Adams H, Jones S, Kane E, Loecke T, Mahaney W, Martina H, Prather C, Robinson T, et al. 2010. Direct and terrestrial vegetation-mediated effects of environmental change on aquatic ecosystem processes. BioScience. 60:590–601.

Bastviken D, Cole J, Pace M, Tranvik L. 2004. Methane emissions from lakes: dependence of lake characteristics, two regional assessments, and a global estimate. Global Biogeochem Cy. 18:1–12.

Bastviken D, Tranvik L, Downing J, Crill P, Enrich-Prast A. 2011. Freshwater methane emissions offset the continental carbon sink. Science. 331:50.

Battin TJ, Luyssaert S, Kaplan LA, Aufdenkampe AK, Richter A, Tranvik LJ. 2009. The boundless carbon cycle. Nat Geosci. 2:598–600.

Baulch HM, Dillon PJ, Maranger R, Schiff SL. 2011. Diffusive and ebullitive transport of methane and nitrous oxide from streams: are bubble-mediated fluxes important? J Geophys Res. 116:G04028.

Bodegom PM, Scholten JCM, Stams AJM. 2004. Direct inhibition of methanogenesis by ferric iron. FEMS Microbiol Ecol. 49:261–268.

Bott TL, Newbold JD, Arscott DB. 2006. Ecosystem metabolism in piedmont streams: reach geomorphology modulates the influence of riparian vegetation. Ecosystems 9:398–421.

Boudreau BP. 2012. The physics of bubbles in surficial, soft, cohesive sediments. Mar Petrol Geol. 38:1–18.

Burrell T, O'Brien J, Graham S. 2014. Riparian shading mitigates stream eutrophication in agricultural catchments. Freshwater Sci. 33:73–84.

Butman D, Raymond PA. 2011. Significant efflux of carbon dioxide from streams and rivers in the United States. Nat Geosci. 4:839–842.

Cole JJ, Prairie Y, Caraco N, McDowell WH, Tranvik LJ, Striegl RG, Duarte CM, Kortelainen P, Downing JA, Middelburg JJ, Melack J. 2007. Plumbing the global carbon cycle: integrating inland waters into the terrestrial carbon budget. Ecosystems. 10:172–185.

Crawford JT, Stanley EH, Spawn SA, Finlay JC, Loken LC, Striegl RG. 2014. Ebullitive methane emissions from oxygenated wetland streams. Glob Change Biol. 20:3408–3422.

Crawford JT, Striegl RG, Wickland KP, Dornblaser MM, Stanley EH. 2013. Emissions of carbon dioxide and methane from a headwater stream network of interior Alaska. J Geophys Res: Biogeosci. 118:1–13.

Denfeld BA, Frey KE, Sobczak WV, Mann PJ, Holmes RM. 2013. Summer CO$_2$ evasion from streams and rivers in the Kolyma River basin, north-east Siberia. Polar Res. 1:1–15.

Fendinger NJ, Adams DD, Glotfelty DE. 1992. The role of gas ebullition in the transport of organic contaminants from sediments. Sci Total Environ. 112:189–201.

Hugelius G, Strauss J, Zubrzycki S, Harden JW, Schuur EAG, Ping C-L, Schirrmeister L, Grosse G, Michaelson GJ, Koven CD, et al. 2014. Estimated stocks of circumpolar permafrost carbon with quantified uncertainty ranges and identified data gaps. Biogeosciences. 11:6573–6593.

Huotari J, Nykänen H, Forsius M, Arvola L. 2013. Effect of catchment characteristics on aquatic carbon export from a boreal catchment and its importance in regional carbon cycling. Glob Change Biol. 19:3607–3620.

Kirschke S, Bousquet P, Ciais P, Saunois M, Canadell JG., Dlugokencky EJ, Bergamaschi P, Bergmann D, Blake DR, Bruhwiler L, et al. 2013. Three decades of global methane sources and sinks. Nat Geosci. 6:813–823.

Kling G, Kipphut G, Miller M. 1992. The flux of CO$_2$ and CH$_4$ from lakes and rivers in arctic Alaska. Hydrobiologia. 240:23–36.

Lundin EJ, Giesler R, Persson A, Thompson MS, Karlsson J. 2013. Integrating carbon emissions from lakes and streams in a subarctic catchment. J Geophys Res: Biogeosci. 118:1–8.

Maeck A, DelSontro T, Mcginnis DF, Fischer H, Flury S, Schmidt M, Fietzek P, Lorke A. 2013. Sediment trapping by dams creates methane emission hot spots. Environ Sci Technol. 47:8130–8137.

Maeck A, Hofmann H, Lorke A. 2014. Pumping methane out of aquatic sediments – ebullition forcing mechanisms in an impounded river. Biogeosciences. 11:2925–2938.

Mattson M, Likens G. 1990. Air pressure and methane fluxes. Nature. 347:718–719.

Molongoski JJ, Klug MJ. 1980. Anaerobic metabolism of particulate organic matter in the sediments of a hypereutrophic lake. Freshwater Biol. 10:507–518.

Overland J, Wang M, Walsh J, Stroeve J. 2014. Future Arctic climate changes: adaptation and mitigation time scales. Earth's Future. 2:1–7.

Pena EA, Slate EH. 2014. gvlma: Global validation of linear models assumptions. R package version 1.0.0.2. http://CRAN.R-project.org/package=gvlma

Raymond PA, Hartmann J, Lauerwald R, Sobek S, McDonald C, Hoover M, Butman D, Striegl R, Mayorga E, Humborg C, et al.

2013. Global carbon dioxide emissions from inland waters. Nature. 503:355–359.

Roden EE, Wetzel RG. 1996. Organic carbon oxidation and suppression of methane production by microbial Fe(III) oxide reduction in vegetated and unvegetated freshwater wetland sediments. Limnol Oceanogr. 41:1733–1748.

Sawakuchi HO, Bastviken D, Sawakuchi AO, Krusche A V, Ballester MVR, Richey JE. 2014. Methane emissions from Amazonian rivers and their contribution to the global methane budget. Glob Change Biol. 20:2829–2840.

Schimel J. 1995. Plant transport and methane production as controls on methane flux from arctic wet meadow tundra. Biogeochemistry. 28:183–200.

Schuur EAG, Abbott BW, Bowden WB, Brovkin V, Camill P, Canadell JG, Chanton JP, Chapin FS, Christensen TR, Ciais P, et al. 2013. Expert assessment of vulnerability of permafrost carbon to climate change. Climatic Change. 119:359–374.

Schuur EAG, Bockheim J, Canadell JG, Euskirchen E, Christopher B, Goryachkin SV, Hagemann S, Kuhry P, Lafleur PM, Lee H, et al. 2008. Vulnerability of permafrost carbon to climate change: implications for the global carbon cycle. BioScience. 58:701–714.

Striegl RG, Dornblaser MM, McDonald CP, Rover JR, Stets EG. 2012. Carbon dioxide and methane emissions from the Yukon River system. Glob Biogeochem Cy. 26:GB0305.

Tarnocai C, Canadell JG, Schuur EAG, Kuhry P, Mazhitova G, Zimov S. 2009. Soil organic carbon pools in the northern circumpolar permafrost region. Glob Biogeochem Cy. 23:GB2023.

van Kesteren W, van Kessel T. 2002. Gas bubble nucleation and growth in cohesive sediments. In: Winterwerp JC, Kranenburg C, editors. Fine sediment dynamics in the marine environment. Amsterdam (Netherlands): Elsevier. p. 329–341.

Varadharajan C, Hemond H. 2012. Time-series analysis of high-resolution ebullition fluxes from a stratified, freshwater lake. J Geophys Res. 117:G02004.

Walter KM, Zimov SA, Chanton JP, Verbyla D, Chapin III FS. 2006. Methane bubbling from Siberian thaw lakes as a positive feedback to climate warming. Nature. 443:71–75.

Wik M, Crill PM, Varner RK, Bastviken D. 2013. Multiyear measurements of ebullitive methane flux from three subarctic lakes. J Geophys Res: Biogeosci. 118:1307–1321.

Wik M, Thornton B, Bastviken D, MacIntyre S, Varner RK, Crill PM. 2014. Energy input is primary controller of methane bubbling in subarctic lakes. Geophys Res Lett. 41:555–560.

Wilcock RJ, Sorrell BK. 2008. Emissions of greenhouse gases CH_4 and N_2O from low-gradient streams in agriculturally developed catchments. Water Air Soil Pollut. 188:155–170.

Yvon-Durocher G, Allen AP, Bastviken D, Conrad R, Gudasz C, St-Pierre A, Thanh-Duc N, Del Giorgio PA. 2014. Methane fluxes show consistent temperature dependence across microbial to ecosystem scales. Nature. 507:488–491.

Yvon-Durocher G, Caffrey JM, Cescatti A, Dossena M, Del Giorgio PA, Gasol JM, Montoya JM, Pumpanen J, Staehr P, Trimmer M, et al. 2012. Reconciling the temperature dependence of respiration across timescales and ecosystem types. Nature. 487:472–476.

Zimov SA, Voropaev Y, Semiletov I, Davidov S, Prosiannikov S, Chapin III FS, Chapin M, Trumbore S, Tyler S. 1997. North Siberian lakes: a methane source fueled by Pleistocene carbon. Science. 800.

Supplementary Material

Supplementary Material is available for download via the Inland Waters website, https://www.fba.org.uk/journals/index.php/IW:

Supplementary Figure S1-2; Supplementary Table S1; Supplementary Methods.